



## Immobilization of hydrothermally produced $\text{TiO}_2$ with different phase composition for photocatalytic degradation of phenol

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### ABSTRACT

Hydrothermally produced  $\text{TiO}_2$  powders with different phase composition (anatase, rutile and mixed phase) were immobilized on glass fibers and tested in the phenol mineralization process. Both  $\text{H}_2\text{O}_2$  and  $\text{O}_2$  were used as oxygen donors, and their performances were compared with those of the same  $\text{TiO}_2$  samples as slurries.

The catalytic properties of the immobilized different crystalline phases, rutile and anatase, show the same trend as the slurry samples: pure rutile displays the highest catalytic efficiency in the presence of  $\text{H}_2\text{O}_2$ , while samples containing anatase improve the photodegradation efficacy with  $\text{O}_2$ . It was suggested that the stability of the photogenerated electron–hole couple allows high activity of rutile in the presence of  $\text{H}_2\text{O}_2$ , while the relevant oxygen chemisorption on anatase causes high catalytic activity in the presence of  $\text{O}_2$ . A four parameters kinetics model shows that both reaction steps, the phenol degradation and the mineralization of the intermediates, are photoactivated by  $\text{TiO}_2$ .

Photoactivity of the coated glass fibers is generally lower than that of slurries, even if their efficiencies are almost comparable when the oxidation is performed by  $\text{H}_2\text{O}_2$ , while much lower when the oxygen donor is  $\text{O}_2$ . As a matter of fact, the morphology of immobilized catalysts shows the presence of chestnut burr aggregates of large rutile crystalline rods on the glass fiber, which are much less compact than the aggregates of small anatase particles. This preserves rutile surface area from the coarsening effects; thus, when rutile is the more active species, as in the presence of  $\text{H}_2\text{O}_2$ , the photocatalytic activity is less affected by immobilization.

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## 1. Introduction

Recent literature reports about several attempts to immobilize  $\text{TiO}_2$  powder on different supporting materials [1–21], when titania is used as photocatalyst for the degradation of hazardous organic compounds in air or water. In fact the simple use of powder catalyst, especially in the nanosized form, has the disadvantage of a difficult sedimentation after the photocatalytic process; in addition  $\text{TiO}_2$  nanoparticles, dispersed in the surrounding, may be hazardous, due to their potential inflammatory and cytotoxic effects [22].

$\text{TiO}_2$  was generally supported on  $\text{SiO}_2$  glass beads [1], rings [2], and reactor tubular walls [3,4]; fiber-glass [5,6]; quartz [7,8]; zeolites [9,10]; perlite [11]; pumice [12]; alumina-based ceramic [13]; stainless steel [14,15]; aluminum [16]; polymeric membranes [17–19], etc. Recently natural fibers began to be used as

suitable supports and the cotton material was conveniently adapted inside the photoreactors [20,21].

In spite of the different  $\text{TiO}_2$  crystalline phases (anatase, rutile and brookite), most of the literature on immobilized catalysts is focused on the use of commercial P25 photocatalyst or of anatase, due to the small average size of the particles (<50 nm); in fact it is expected that the smaller the particle size the higher the surface area suitable for the catalytic activity. No attention, until lately, was directed to the possibility that other properties beside the surface area, like crystal shape and the oxygen source, could be influent.

Very recently we demonstrated that the catalytic activity of  $\text{TiO}_2$  in the degradation of phenol ( $\text{PhOH}$ ) in water depends on both the average particle size and the type of oxygen donor ( $\text{O}_2$  or  $\text{H}_2\text{O}_2$ ). Specifically, when the oxygen donor was  $\text{H}_2\text{O}_2$ , the large crystals of pure rutile or rutile-rich mixtures showed the highest activity; oppositely, when using  $\text{O}_2$ , anatase and anatase-rich mixtures are the most active [23,24]. This behavior was attributed to the slow recombination rate of the electron–hole couple in rutile large nanocrystals, which controls the catalytic mechanism when the

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oxidant is  $\text{H}_2\text{O}_2$ . On the contrary, the easy oxygen chemisorption at the surface of the small anatase particles controls the mechanism when the oxygen donor is  $\text{O}_2$ .

The present paper reports about the immobilization on glass fibers, of a number of hydrothermal titania samples with different phase composition and, as a consequence, with different average particle size. Catalysts were characterized both as pure and supported powders. The catalytic activity of the samples as slurry was compared with that of the immobilized ones in the PhOH degradation reaction, using both  $\text{H}_2\text{O}_2$  and  $\text{O}_2$ . The aim was to investigate how much the immobilization process affects the catalytic activity, and to verify whether the trend of the catalytic activity observed in slurry samples keeps the same in supported samples.

## 2. Experimental

### 2.1. Synthesis of $\text{TiO}_2$ and immobilization on glass fibers

Nanocrystalline  $\text{TiO}_2$  was obtained by hydrothermal synthesis according to the previously reported procedure [23] by reacting aqueous solutions of  $\text{TiOCl}_2$  (Aldrich, 99%) and  $\text{NH}_3$  (Fluka, >25 wt%) in a teflon lined autoclave (Parr, model 4768Q). The autoclave was heated at 2.67 °C/min to a temperature of 30 °C below the set up temperature and then at 0.75 °C/min up to 220 °C.

The suspension was aged for 2 h at this temperature. The autoclave was then cooled in air outside the oven until the temperature dropped below 70 °C. After decantation,  $\text{TiO}_2$  powder was recovered from the autoclave and washed first with water, then with ammonia solution and finally with acetone. Titanium dioxide nanoparticles were recovered by filtration and dried under vacuum at room temperature. The phase composition was determined by X-ray diffraction (XRD) patterns (see Section 2.2).

$\text{TiO}_2$  samples with different phase composition, labeled as Rnn, where nn represents the weight% of rutile phase (Table 1) were selected for the immobilization on glass fiber (glass fiber filter Whatman GF/A).  $\text{TiO}_2$  powder (1.0 g) was suspended by sonication in 13.2 mL of water. While stirring, polyethylene glycol (PEG 2000) aqueous solution (5.0 g, 50 wt%) and few drops of Triton X-100 were added sequentially in order to obtain the suitable viscosity for a good and uniform coating adhesion. The mixture was homogeneously pasted on the glass fiber material (size 7 cm × 20 cm) with a brush and left to dry in an oven at 120 °C for 30 min. Successively, the coating was calcined at 400 °C in air for 30 min to fully remove the organic content. The final amount of  $\text{TiO}_2$  deposited on glass fiber was  $5.5 \pm 0.5 \text{ g cm}^{-2}$ . Finally, such immobilized photocatalyst was inserted into the reactor for the photocatalytic tests.

Hereafter,  $\text{TiO}_2$  coated glass fiber were labeled as Rnn (i), where nn stand for the weight% of rutile phase and (i) indicates the immobilization on the fiber.

**Table 1**

Phase composition, BET specific surface area and XRD particle size of hydrothermal  $\text{TiO}_2$  powder samples.

	Crystal phase composition		BET surface area ( $\text{m}^2 \text{ g}^{-1}$ )	XRD particle size (nm)	
	Anatase (wt%)	Rutile (wt%)		Anatase d(1 0 1)	Rutile d(1 1 0)
R100	0	100	20	–	58
R70	30	70	46	14	68
R57	43	57	64	15	31
R0	100	0	91	15	–

### 2.2. $\text{TiO}_2$ characterization

The X-ray diffraction patterns of the  $\text{TiO}_2$  powders, collected with a Bruker D8 Advance diffractometer ( $\text{Cu K}\alpha$  radiation) operating in the range  $20\text{--}40^\circ 2\theta$  ( $2\theta$  step 0.020°, counting time 2 s per step), were used to determine the crystalline phase and the average crystallite size.

The diffraction peaks were indexed as pure rutile  $\text{TiO}_2$  (JCPDS, no. 21-1276), and pure anatase (JCPDS, no. 21-1272). The phase contents were estimated using the following equation [25]:

$$f_A = \frac{1}{(1 + 1/KI_R/I_A)} \quad (1)$$

where  $f_A$  is the fraction of the anatase phase in the powder;  $K = 0.79$  when  $f_A > 0.2$  and  $K = 0.68$  when  $f_A < 0.2$ ;  $I_R$  and  $I_A$  are the integrated intensities of the rutile 1 1 0 and anatase 1 0 1 peaks, respectively.

The average crystallite size of  $\text{TiO}_2$  was estimated from the broadening of the XRD peaks 1 1 0 for rutile and 1 0 1 for anatase by means of the Scherrer equation.

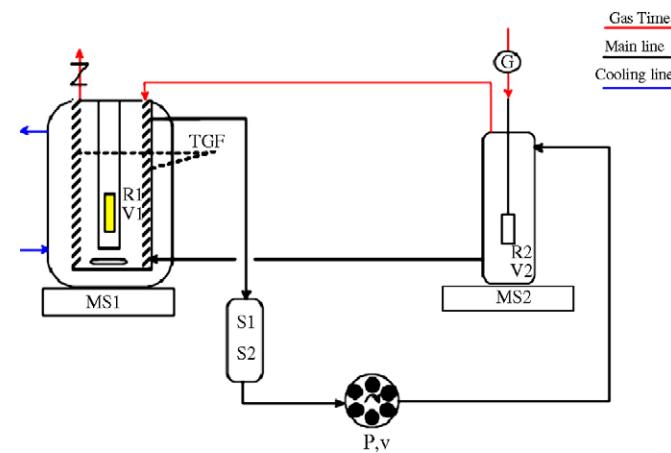
$\text{TiO}_2$  powder specific surface area (BET method) was determined by nitrogen physisorption (Micromeritics, ASAP 2010), after evacuation at 200 °C for 12 h.

Scanning electron microscopy (SEM) measurements were performed by a LEO 1450VP instrument both on powders and on supported catalysts.

High-resolution transmission electron microscopy (HRTEM) and electron diffraction (ED) were performed using a Jeol 3010 apparatus operated at 300 kV with a high-resolution pole piece (0.17 nm point-to-point resolution) and equipped with a Gatan slow-scan 794 CCD camera. The powders were suspended in isopropanol, and a 5  $\mu\text{L}$  drop of this suspension was deposited on a holey carbon film supported on 3-mm copper grid for TEM investigation.

### 2.3. Photocatalytic experiments

Photodegradation experiments were carried out in a 600 mL discontinuous batch reactor with an external cooling jacket, and equipped with a UV 125 W Hg high-pressure lamp, placed in a coaxial quartz cylinder. No optical filter was adopted. The scheme of the photocatalytic apparatus is reported in Fig. 1. The experiments were performed using both  $\text{TiO}_2$  coated glass fibers and  $\text{TiO}_2$  slurries as catalysts in water.

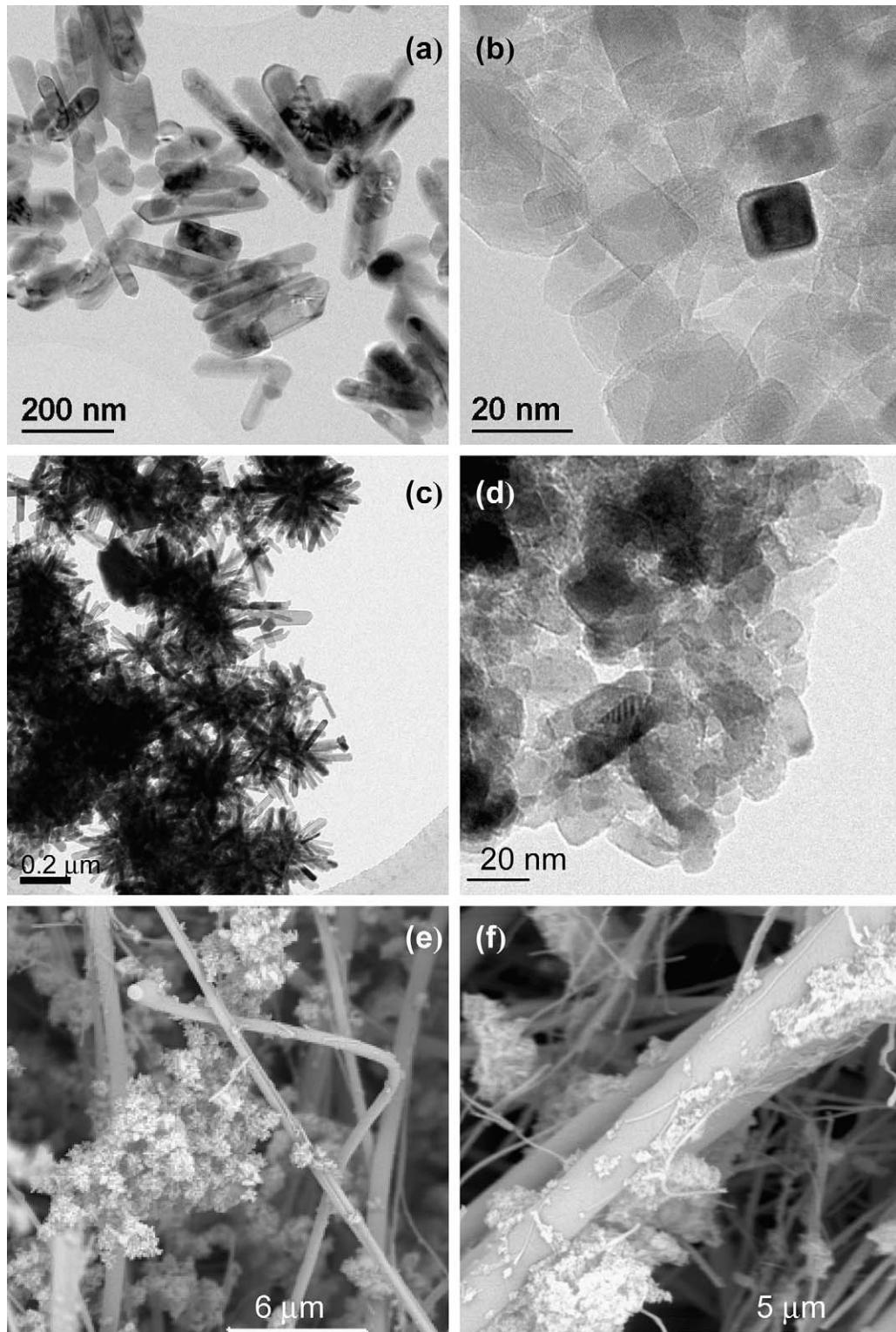


**Fig. 1.** Scheme of the photocatalytic apparatus used for the catalytic test. R1: photoreactor (volume, V1); R2: saturator (volume, V2); P: peristaltic pump (volumetric flux, v); S1, S2: sensors (pH, DO: dissolved oxygen); MS1, MS2: magnetic stirrers; G: gas flow meter; TGF:  $\text{TiO}_2$  coated glass fiber.

In the former case the glass fiber was placed inside the batch reactor, stucked to the inner cylindrical walls at 4 cm from the coaxial UV lamp. PhOH solution (600 mL containing  $121 \pm 2 \text{ mg L}^{-1}$  of PhOH,  $93 \pm 2 \text{ mg L}^{-1}$  as C) was recirculated by a peristaltic pump with a speed of  $14 \text{ mL s}^{-1}$ . The temperature was kept  $25 \pm 2 \text{ }^\circ\text{C}$ ; pH and dissolved oxygen content of the suspension were monitored by on-line sensors. In the case of the slurry,  $\text{TiO}_2$  powder ( $150 \pm 5 \text{ mg; } 0.25 \text{ g L}^{-1}$ )

was suspended by sonication in 600 mL of water containing the same concentration of PhOH and recirculated with the same speed and at the same temperature as in the experiments on immobilized samples. It was verified that the catalyst loading in the range  $0.2\text{--}1.4 \text{ g L}^{-1}$  does not affect the kinetics parameters of PhOH mineralization.

Photodegradation was performed by using  $\text{H}_2\text{O}_2$  or  $\text{O}_2$  as oxygen donors, and operating either with slurries or with



**Fig. 2.** TEM micrograph of (a) R100 and (b) R0 hydrothermal powders; (c) R100 and (d) R0 hydrothermal powders annealed at  $400 \text{ }^\circ\text{C}$  in air; SEM micrographs (HV = 15.2 kV) of (e) R100 (i) (magnification 26.67k $\times$ ) and (f) R0 (i) (magnification 19.67k $\times$ ) coated glass fibers.

immobilized samples of photocatalysts, in order to compare the different experimental procedures. In the experiments performed with  $\text{H}_2\text{O}_2$  (partly heterogeneous oxidation), stoichiometric amounts of aqueous  $\text{H}_2\text{O}_2$  (35 wt%) were added ( $\text{H}_2\text{O}_2/\text{PhOH} = 14 - \text{mol/mol}$ ) and the suspensions or the solutions, when using the immobilized photocatalysts, were circulated in the dark for 30 min before turning on the UV source. When the oxidations were carried out in the presence of  $\text{O}_2$ , and in the absence of any  $\text{H}_2\text{O}_2$  (entirely heterogeneous oxidation), the suspensions or the solutions were circulated in the dark and saturated in an on-line chamber by continuously bubbling oxygen ( $100 \text{ mL min}^{-1}$ ); the excess of gas was eliminated through the non-return check valve. The UV source was turned on when the oxygen content in the suspension became maximum and constant (in about 10 min). The reaction was performed by keeping the oxygen feed constant. In both cases reference experiments were carried out in the absence of  $\text{TiO}_2$  (blank), as well as by using commercial Degussa P25 titanium dioxide as photocatalyst. The kinetics of photoinduced PhOH degradation was monitored until the complete mineralization of the substrate, by withdrawing suspension aliquots (6 mL) of the reaction solution at regular intervals and analyzing them for the total organic carbon (TOC) by a Shimadzu TOC-V CSH analyzer. In the case of the slurry,  $\text{TiO}_2$  powder was separated by centrifugation before TOC analysis.

### 3. Results and discussion

#### 3.1. Morphological characterization of catalysts

TEM micrographs of R100 and R0 powders are reported in Fig. 2(a) and (b), respectively. Rutile particles show elongated shape with a width of about 60 nm and an aspect ratio  $\alpha$  ( $\alpha = \text{length}/\text{width}$ ) of about 5. Anatase particles are small and almost cubic (5–15 nm) with  $\alpha = 1-3$ .

Sizes and shapes of rutile and anatase crystals in mixed phase samples (not shown in figure) reflect those observed in the corresponding pure phases. The average crystal size calculated from the broadening of XRD 1 0 1 anatase peak (Table 1) is in agreement with TEM results. As the rutile crystals grow along the  $c$ -axis, the particle size calculated from the broadening of XRD 1 1 0 peaks corresponds to the width of the crystal.

After further annealing in air at  $400^\circ\text{C}$ , the temperature of glass fiber calcination, both R100 and R0 powders undergo aggregation. R100 forms chestnut burr aggregates of rutile prismatic rods (Fig. 2(c)), while R0 more compact aggregates of square ended anatase nanoparticles (Fig. 2(d)).

SEM micrographs of  $\text{TiO}_2$  coated glass fibers R100 (i) and R0 (i) show the presence of immobilized aggregates (Fig. 2(e) and (f))

whose morphology is reminiscent of that observed in the corresponding powders.

#### 3.2. Photocatalytic activity

The photooxidation mechanism, as suggested by several authors [26,27], implies that electron–hole pairs are generated into  $\text{TiO}_2$  by supra-band gap UV irradiation.

Free holes ( $\text{h}^+$ ) are trapped by the organic molecule on the surface, giving direct photogeneration of organic radicals:



alternatively they are fixed to terminal lattice oxide ions giving  $\text{O}^-$  and  $\cdot\text{OH}$  radicals:



$\text{O}^-$  and  $\cdot\text{OH}$  species react with the organic substrates, generating the corresponding radicals [28]:

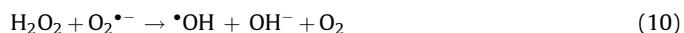


which are further oxidized to other intermediates [29–31] until complete mineralization to  $\text{CO}_2$  and  $\text{H}_2\text{O}$ .

Photogenerated conduction band electrons ( $e^-$ ) may interact with the oxidant ( $\text{H}_2\text{O}_2$  or  $\text{O}_2$ ) according to the following reactions [26,27,32,33]:



where  $\text{O}_2\cdot^-$  may further generate  $\text{H}_2\text{O}_2$ :



and successively,  $\cdot\text{OH}$  radicals via reaction (5).

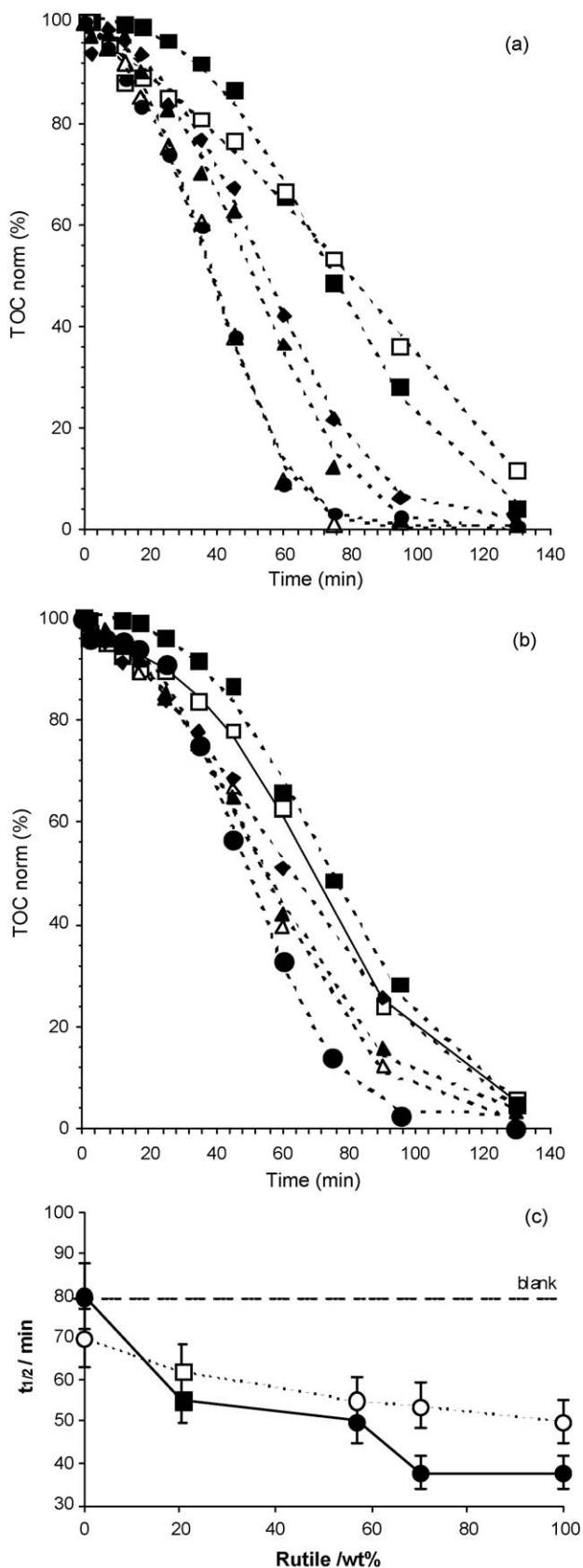
Besides, when using  $\text{H}_2\text{O}_2$ , direct photolysis of peroxide may also produce hydroxyl radicals [34,35]:



The photocatalytic activity of  $\text{TiO}_2$  with different phase composition (Table 1) either immobilized on glass fiber as slurry in water, was measured in the PhOH mineralization, using  $\text{H}_2\text{O}_2$  or

**Table 2**  
Photocatalytic parameters of slurry and glass fiber immobilized  $\text{TiO}_2$  samples.

Samples	Rutile fraction (wt%)	Photocatalytic parameters			
		$\text{H}_2\text{O}_2$	$\text{O}_2$	$(dC/dt)_{\text{max}}$ ( $\text{mg L}^{-1} \text{min}^{-1}$ )	$t_{1/2}$ (min)
R100	100	2.00	0.155	2.00	349
R70	70	2.09	0.379	1.82	153
R57	57	1.61	0.457	1.36	134
R0	0	0.84	0.459	0.84	128
P25	20	1.50	0.682	1.50	95
Blank	–	1.27	0.075	1.27	740
R100 (i)	100	1.82	0.065	1.82	658
R70 (i)	70	1.36	0.058	1.36	639
R57 (i)	57	1.27	0.192	1.27	399
R0 (i)	0	1.21	0.210	1.21	472
P25 (i)	20	0.98	0.220	0.98	524



**Fig. 3.** Mineralization curves of phenol (given as TOC%) in the presence of the catalysts (a) in slurry and (b) immobilized on glass fiber using  $\text{H}_2\text{O}_2$  as oxygen donor: (■) blank; (□) RO; (△) R70; (▲) R57; (●) R100; (◆) P25. Dashed lines are the best fit to the experimental data according to Eq. (12). (c) Half transformation time ( $t_{1/2}$ ) of TOC against rutile fraction (wt%) for hydrothermal powder (●, ○) and P25 (■, □) samples in the slurry (full dots) and immobilized on glass fiber (empty dots). 100% TOC at  $t = 0$  corresponds to phenol amount before UV irradiation.

$\text{O}_2$  as oxygen donors under UV irradiation. The catalyst activity was evaluated by reproducing the experimental data (TOC vs. time) with the function

$$C(t) = C_0 - A \int_0^x \exp\left(-\frac{(t - t_0)^2}{s}\right) dt \quad (12)$$

where  $C(t)$  is the TOC amount measured at the  $t$  time and  $C_0, A, t_0, s$  are fitting parameters ( $C_0$  is the total organic carbon in solution after 30 min in the dark before the photocatalytic PhOH degradation and  $t_0$  the time at which the mineralization of the organic substrate starts). The maximum degradation rate  $(dC/dt)_{\max}$ , which corresponds to the maximum slope point of the fitted curves, and the half degradation time  $t_{1/2}$ , were taken as representative parameters of the kinetic behavior of the different samples. Both  $(dC/dt)_{\max}$  and  $t_{1/2}$  values were calculated from Eq. (12) after optimization of the fitting parameters and were normalized to the same mass (150 mg) of  $\text{TiO}_2$ .

In all the performed experiments, during the initial phase of recirculation in the dark (30 min), the loaded PhOH concentration decreased of about  $5.0 \pm 1.5\%$  in presence of both  $\text{H}_2\text{O}_2$  or  $\text{O}_2$  oxygen donors.

The experimental data related to TOC disappearance and the fitted curves of PhOH mineralization in the presence of  $\text{H}_2\text{O}_2$ , for both  $\text{TiO}_2$  coated glass fiber and slurry, are reported in Table 2 and Fig. 3(a) and (b), where 100% TOC at  $t = 0$  is the PhOH concentration measured just before switching on UV irradiation.

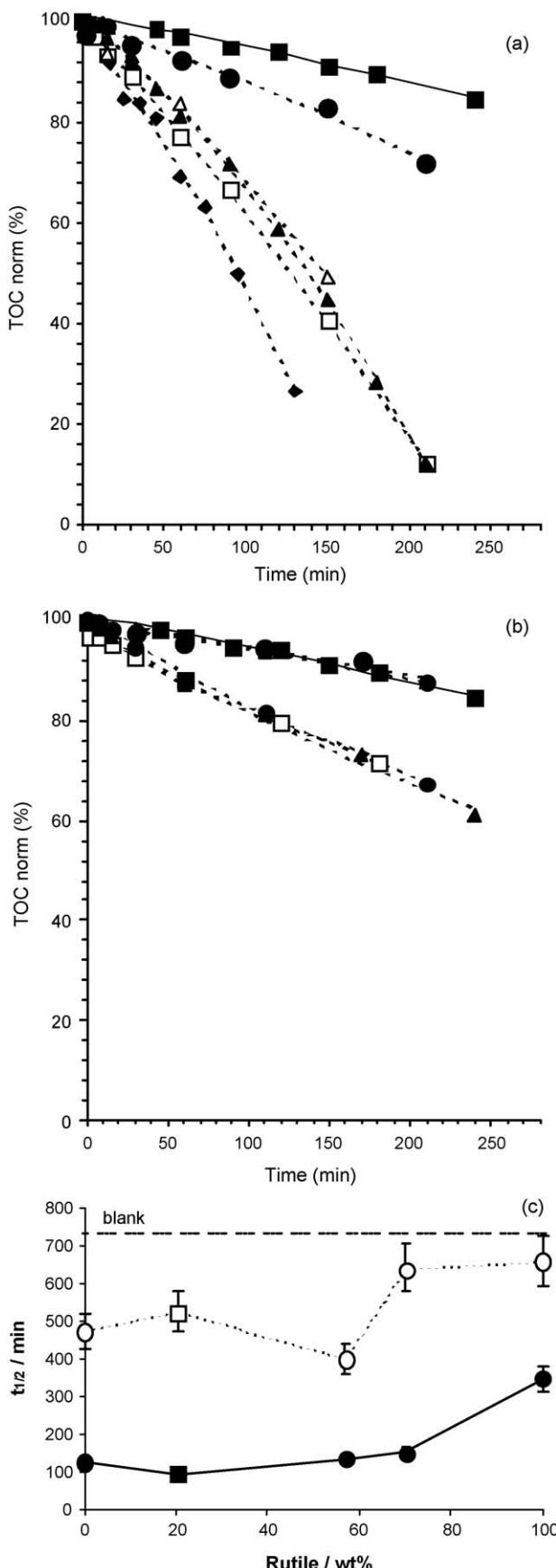
The trend is similar for both types of samples:  $t_{1/2}$  value decreases and  $(dC/dt)_{\max}$  increases by increasing the rutile content and/or the particle size, as already widely discussed in analogous photocatalytic experiments [23,24]. The highest photocatalytic activity was observed for pure rutile, R100, having large prismatic crystals with high aspect ratio  $\alpha$ , while pure anatase, R0, exhibits the lowest activity, of the same order than the blank test performed in the absence of  $\text{TiO}_2$ . These results confirm the relevance of the crystalline rutile nanorods in increasing the photocatalytic activity. Well-crystallized and well faceted nanocrystals involve relatively low concentration of defects, both in the bulk and on the surface, avoiding recombination of the photoexcited electron–hole pairs [36–38].

Accordingly the photoactivity of immobilized  $\text{TiO}_2$  when using  $\text{H}_2\text{O}_2$  as oxidant is only slightly lower than that of  $\text{TiO}_2$  powder oxidized as a slurry (Fig. 3(c)).

If the PhOH mineralization was performed in the presence of  $\text{O}_2$  as oxygen donor, the photocatalytic activity is generally lower than in the case of  $\text{H}_2\text{O}_2$ . Besides the photoactivity of  $\text{TiO}_2$  coated glass fiber is significantly lower than that of  $\text{TiO}_2$  powder in the slurry.

Also in this case, however, the photoactivity trend is similar for both immobilized and slurry samples (Fig. 4(a) and (b)): the values of  $t_{1/2}$  decrease and those of  $(dC/dt)_{\max}$  increase with the increase of the anatase amount (Table 2). Pure rutile shows the lowest photoactivity. These results confirm that the presence of anatase is essential to improve the photocatalytic activity when  $\text{O}_2$  is the oxygen donor [24] and that small differences exist between pure anatase and mixed phase samples. The higher photoactivity of anatase in the presence of  $\text{O}_2$  can be related to its high ability to chemisorb  $\text{O}_2$  as  $\text{O}_2^-$ ; this extends the independent lifetime of electrons and holes, decreasing the  $e^- - h^+$  recombination rate also in small particles [24].

In the blank, without the catalyst, the PhOH mineralization rate is very low, thus indicating a minor and negligible contribution of direct PhOH photolysis to the overall process.



**Fig. 4.** Mineralization curves of phenol (given as TOC%) in the presence of the catalysts (a) in slurry and (b) immobilized on glass fiber using  $O_2$  as oxygen donor: (■) blank; (□) R0; (Δ) R70; (▲) R57; (●) R100; (◆) P25. Dashed lines are the best

The photocatalytic data allow to conclude that the reaction mechanism of immobilized and slurry catalysts is the same and, therefore, the stabilization of the photogenerated electron–hole pairs is the determining step. In particular, in the case of oxidation with  $H_2O_2$  the hindering of the electron–hole recombination is related to the crystal size and shape of particles [23]. Instead, when  $O_2$  is the oxidant, the  $e^-$ – $h^+$  separation depends on the relevant chemisorption of the electron scavenger  $O_2^-$  at the surface of catalyst nanoparticles [24].

### 3.3. Kinetics of PhOH photodegradation in the presence of $H_2O_2$

In order to obtain deeper insight into the catalytic mechanism of the PhOH photodegradation catalyzed both by  $TiO_2$  coated glass fibers and by powder slurry catalysts, the four parameter kinetic model proposed by Bellobono and co-workers for the partly heterogeneous oxidation, such as that given when operating in the presence of hydrogen peroxide as oxygen donor, was employed [39].

According to this model, it was assumed that the PhOH mineralization to  $CO_2$  occurs through a single intermediate I, which contains the contribution of all intermediates:



A system of first order differential equations describes the variation of PhOH, I and  $CO_2$  concentrations [39]:

$$\left( \frac{dC}{dt} \right) = -\frac{k_1 K_1 C}{1 + K_1 C + K_2 C_I} \quad (14)$$

$$\left( \frac{dC_I}{dt} \right) = \frac{k_1 K_1 C k_2 K_2 C_I}{1 + K_1 C + K_2 C_I} \quad (15)$$

$$\left( \frac{dC_{CO_2}}{dt} \right) = \frac{k_1 K_1 C_I}{1 + K_1 C + K_2 C_I} \quad (16)$$

where  $C$ ,  $C_I$  and  $C_{CO_2}$  are the concentrations of PhOH, I and  $CO_2$ ;  $K_1$  and  $K_2$  are the apparent chemisorption constants related to the competitive absorption of, respectively, PhOH and I onto  $TiO_2$  surface;  $k_1$  and  $k_2$  are the kinetic constants of PhOH and I degradation.

When the photodegradation begins,  $C_I$  is negligible with respect to  $C$ , and the kinetic equations can be approximated to:

$$\left( \frac{dC}{dt} \right) = -\frac{k_1 K_1 C}{1 + K_1 C} \quad (17)$$

At the beginning of the process, when the term  $K_1 C > 1$  ( $C_{(0)} = 1.53 \times 10^{-3}$  M and the order of magnitude of  $K_1$  reported for PhOH degradation is  $\approx 10^7$  M $^{-1}$  [39]), Eq. (17) can be further simplified to a pseudo-zero-order kinetic equation, where  $k_{(0)} = k_1$  is the pseudo-zero-order kinetic constant:

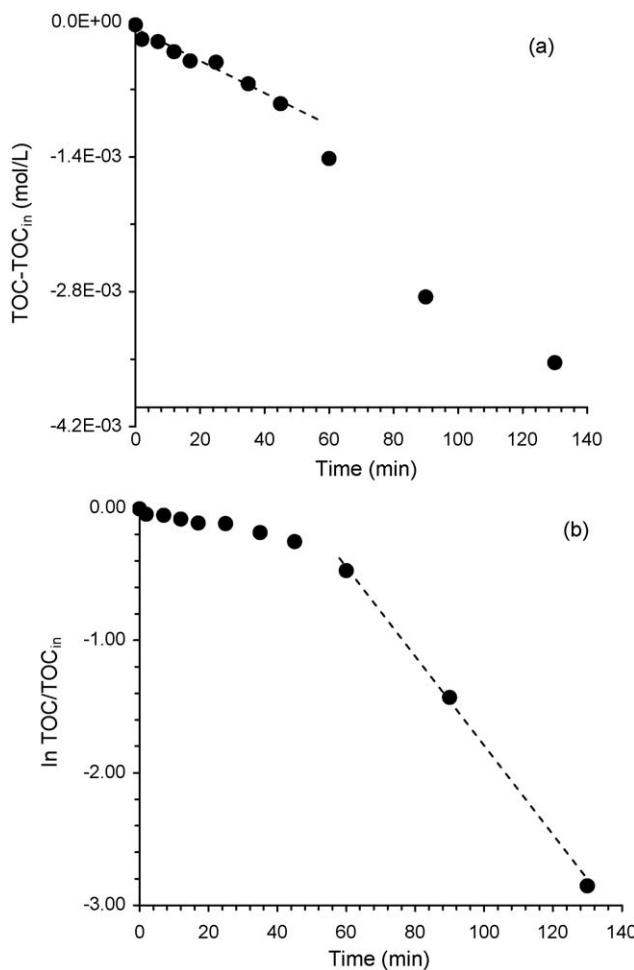
$$\left( \frac{dC}{dt} \right) \approx -k_{(0)} \quad (18)$$

Integrating between  $t = 0$  and  $t$  it results

$$C - C_{(0)} \approx -k_{(0)} t \quad (19)$$

Considering that at the very early stages of reaction the total organic carbon concentration may be assumed  $C_{TOC} \approx 6 C$ , as PhOH contains six carbon atoms, and the very small concentration of intermediates is still given by hydroxylated compounds with the

fit to the experimental data according to Eq. (12). (c) Half transformation time ( $t_{1/2}$ ) of TOC against rutile fraction (wt%) for hydrothermal powder (●, ○) and P25 (■, □) samples in the slurry (full dots) and immobilized on glass fiber (empty dots). 100% TOC at  $t = 0$  corresponds to phenol amount before UV irradiation.



**Fig. 5.** Plots of (a)  $(C_{TOC} - C_{TOC(0)})$  and (b)  $\ln(C_{TOC}/C_{TOC(0)})$  vs. reaction time calculated from a typical experimental TOC vs. time curve (sample R0 (i)) for the phenol photomineralization in the presence of  $H_2O_2$  as oxygen donor, according to Eqs. (20) and (24).

aromatic ring, Eq. (19) can be written as:

$$C_{TOC} - C_{TOC(0)} \approx -k_{(0)} t \quad (20)$$

by expressing the PhOH concentration in terms of carbon concentration.

On the other hand, at a time  $t^*$ , when  $C \ll C_I$ , the kinetic Eq. (15) can be simplified to:

$$\left( \frac{dC_I}{dt} \right) = - \frac{k_2 K_2 C_I}{1 + K_2 C_I} \quad (21)$$

When  $C_I$  is very small, Eq. (21) becomes a pseudo-first-order kinetic equation, where  $k_{(1)} = k_2 K_2$  is the pseudo-first-order kinetic constant:

$$\left( \frac{dC_I}{dt} \right) \approx -k_{(1)} C_I \quad (22)$$

By integrating between  $t^*$  and  $t$  we obtain:

$$\ln \frac{C_I}{C_{I(t^*)}} \approx -k_{(1)} C_I \quad (23)$$

By considering that when  $C \ll C_I$ ,  $C_I \approx C_{TOC}$  and that  $C_{I(t^*)} \approx C_{TOC(0)}$ , Eq. (23) can be rewritten as:

$$\ln \frac{C_{TOC}}{C_{TOC(0)}} \approx -k_{(1)} C_I \quad (24)$$

**Table 3**

Kinetic constants  $k_{(0)}$  and  $k_{(1)}$  of  $TiO_2$  coated glass fiber and powder slurry for the photomineralization of phenol in the presence of stoichiometric hydrogen peroxide.

Sample	$k_{(0)}$ (mol min <sup>-1</sup> )	$k_{(1)}$ (min <sup>-1</sup> )
R100	9.40E-05	1.82E-02
R70	8.45E-05	1.82E-02
R57	7.39E-05	1.39E-02
R0	4.21E-05	8.66E-03
P25	6.47E-05	1.26E-02
Blank	8.23E-06	9.37E-03
R100 (i)	1.36E-05	1.39E-02
R70 (i)	1.15E-05	1.28E-02
R57 (i)	1.18E-05	1.26E-02
R0 (i)	8.69E-06	9.90E-03
P25 (i)	9.84E-06	1.12E-02

In Fig. 5(a) and (b) the plots of, respectively  $(C_{TOC} - C_{TOC(0)})$  and  $\ln(C_{TOC}/C_{TOC(0)})$  vs. reaction time, are reported, as calculated for a typical TOC vs. time curve (sample R0 in Fig. 3(a)).

It results that at the beginning of PhOH degradation, the kinetics is of apparent zero order (Fig. 5(a)) while at the end of the process the kinetics approximate a first order reaction. According to Eqs. (19) and (24),  $k_{(0)}$  and  $k_{(1)}$  can be calculated from the angular coefficients of the straight lines fitting the experimental points.

The values of  $k_{(0)}$  and  $k_{(1)}$  for all  $TiO_2$  samples, calculated for both  $TiO_2$  coated glass fiber and powder slurry samples from the experimental photodegradation data, are reported in (Table 3).

The fitting of the experimental data by the use of the equations above discussed, demonstrates that the kinetic hypotheses are correctly formulated. Moreover the results show that both  $k_{(0)}$  and  $k_{(1)}$  values increase with the amount of rutile, their trend resulting parallel to that of  $(dC/dt)_{max}$  and  $t_{1/2}$  (Table 2). This suggests that the photocatalytic activity of  $TiO_2$ , in the presence of hydrogen peroxide as oxygen donor, is depending on the rate of electron-hole recombination both during the initial PhOH photodegradation and during the mineralization of the intermediate compounds.

#### 4. Conclusions

The paper reports on the photoactivity of hydrothermal  $TiO_2$  with different phase compositions, immobilized on glass fiber, used as a catalyst in the PhOH oxidative degradation by  $H_2O_2$  and  $O_2$ .

The photocatalytic tests show that immobilized catalysts are active both in partly homogenous (use of  $H_2O_2$  as oxygen donor) and entirely heterogeneous conditions (use of  $O_2$  as oxygen donor). The activity is generally lower for coated glass fibers, if compared with the performances of the same  $TiO_2$  samples in slurries; when the oxidation employs  $H_2O_2$  the reactivity of the immobilized catalyst is comparable (about 1.5 lower) to that of the slurry  $TiO_2$ , while it is much lower when the oxygen donor is  $O_2$  (4–5 times).

Nevertheless, the immobilization process does not modify the specific catalytic properties of rutile and anatase, investigated in our previous papers [23,24]. In partly homogeneous phase, being  $H_2O_2$  the oxygen donor, immobilized pure rutile displays the highest catalytic efficiency, in comparison with mixed phase (anatase/rutile) samples and with pure anatase, which show the lowest photocatalytic activity. In entirely heterogeneous phase, being  $O_2$  the oxygen donor, immobilized anatase, both in pure and mixed phase, is essential to improve the photodegradation, while pure rutile displays the lowest catalytic activity.

Analogously to the slurry catalytic experiments, the very high catalytic activity of rutile when  $H_2O_2$  is used, can be explained by the lower tendency of UV generated electrons and holes to recombine in the high crystalline rods of rutile with high aspect ratio when compared to the smaller nanosized particles of anatase (5–15 nm).

Similarly, the higher activity of anatase when the oxygen donor is  $O_2$  can be related to the easy chemisorption of oxygen as  $O_2^-$  at the anatase surface [24], acting as an effective electron scavenger. This increases the independent lifetime of electrons and holes, decreasing the electron–hole recombination rate.

The morphology of the hydrothermal  $TiO_2$  nanocrystals observed by electron microscopy explains the different catalytic efficiency of immobilized  $TiO_2$  with respect to powder slurry. The large chestnut burr aggregates of rutile crystalline rods do not form compact aggregates when coated on the glass fiber unlike the small anatase particles do. This maintains a higher free surface area of the supported catalyst, suitable both to UV irradiation and to the adsorption/reaction of PhOH substrate and the oxidative agents. Thus when rutile is the more active species, in the presence of  $H_2O_2$  as catalyst, the reduction of the activity due to immobilization is low and the performance of the coated glass fiber comparable with that of powder slurry. When anatase is the more active species, the relevant particle aggregation decreases the chemisorption of oxygen at the surface and leads to a large decrease of activity.

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## References

- [1] M. Tashibi, C.R. Ngah, N. Aziz, A. Mansor, A.Z. Abdullah, L.K. Teong, A.R. Mohamed, Ind. Eng. Chem. Res. 46 (2007) 9006–9014.
- [2] J. Fernandez, J. Kiwi, J. Baeza, J. Freer, C. Lizama, H.D. Mansilla, Appl. Catal. B 48 (2004) 205–211.
- [3] J.C. Lee, M.S. Kim, B.W. Kim, Water Res. 36 (2002) 1776–1782.
- [4] M.D. Nikolaki, D. Malamis, S.G. Poulopoulos, C.J. Philippopoulos, J. Hazard. Mater. 137 (2006) 1189–1196.
- [5] S. Horikoshi, N. Watanabe, H. Onishi, H. Hikada, N. Serpone, Appl. Catal. B 37 (2002) 117–129.
- [6] M.K. Aminian, N. Taghavinia, A. Irajizad, S.M. Mahdavi, J. Phys. Chem. C 111 (2007) 9794–9798.
- [7] I.N. Martyanov, N.J. Klabunde, J. Catal. 225 (2004) 408–416.
- [8] M.G. Antoniou, D.D. Dionysiou, Catal. Today 124 (2007) 215–223.
- [9] Y. Xu, C.H. Langford, J. Phys. Chem. B 101 (1997) 3115–3121.
- [10] F. Li, S. Sun, Y. Jiang, M. Xia, M. Sun, B. Xue, J. Hazard. Mater. 152 (2008) 1037–1044.
- [11] S.N. Hosseini, S.M. Borghei, M. Vossoughi, N. Taghavinia, Appl. Catal. B 74 (2007) 53–62.
- [12] S.K. Kansal, M. Singh, D. Sud, J. Hazard. Mater. 153 (2008) 412–417.
- [13] T.H. Xie, J. Lin, J. Phys. Chem. C 111 (2007) 9968–9974.
- [14] J. Shang, W. Li, Y. Zhu, Mol. Catal. A 202 (2003) 187–195.
- [15] N. Kieda, T. Tokuhisa, J. Ceram. Soc. Jpn. 114 (2006) 42–45.
- [16] H. Chen, S.W. Lee, T.H. Kim, B.Y. Hur, J. Eur. Ceram. Soc. 26 (2006) 2231–2239.
- [17] I.R. Bellobono, R. Barni, F. Gianturco, J. Membr. Sci. 102 (1995) 139–147.
- [18] F. Ascari, I.R. Bellobono, P.M. Tozzi, Fresen. Environ. Bull. 12 (2003) 1195–1201.
- [19] I.R. Bellobono, F. Ascari, C. Lagrasta, Fresen. Environ. Bull. 12 (2003) 1536–1544.
- [20] Z. Liuxue, W. Xiulian, L. Peng, S. Zhixing, Surf. Coat. Technol. 201 (2007) 7607–7614.
- [21] B. Tryba, J. Hazard. Mater. 151 (2008) 623–627.
- [22] L. Reijnders, J. Hazard. Mater. 152 (2008) 440–445.
- [23] A. Testino, I.R. Bellobono, V. Buscaglia, C. Canevali, M. D'Arienzo, S. Polizzi, R. Scotti, F. Morazzoni, J. Am. Chem. Soc. 129 (2007) 3564–3575.
- [24] R. Scotti, I.R. Bellobono, C. Canevali, C. Cannas, M. Catti, M. D'Arienzo, A. Musinu, S. Polizzi, M. Sommariva, A. Testino, F. Morazzoni, Chem. Mater. 20 (2008) 4051–4061.
- [25] K.S. Mazdiyasni, Ceram. Int. 8 (1982) 42–56.
- [26] A. Sclafani, J.M. Herrmann, J. Phys. Chem. 100 (1996) 13655–13661.
- [27] P. Salvador, J. Phys. Chem. C 111 (2007) 17038–17043.
- [28] U. Diebold, Surf. Sci. Rep. 48 (2003) 53–229.
- [29] P. Davit, G. Martra, S. Coluccia, J. Jpn. Petrol. Inst. 47 (2004) 359–376.
- [30] J. Theurich, M. Lindner, D.W. Bahneemann, Langmuir 12 (1996) 6368–6376.
- [31] R. Vinu, G. Madras, Environ. Sci. Technol. 42 (2008) 913–919.
- [32] M. Ampo, T. Shima, Y. Kubokawa, Chem. Lett. 12 (1985) 1799–1801.
- [33] K. Okamoto, Y. Yakamoto, H. Tanaka, M. Tanaka, M.A. Itaya, Bull. Chem. Soc. Jpn. 58 (1985) 2015–2016.
- [34] E. Selli, I.R. Bellobono, M.L. Raimondi, Angew. Makromol. Chem. 196 (1992) 169–177.
- [35] J.H. Baxendale, J.A. Wilson, Trans. Faraday Soc. 53 (1957) 344–355.
- [36] T. Ohno, K. Sarukawa, M. Matsumura, New J. Chem. 26 (2002) 1167–1170.
- [37] Y. Wang, L. Zhang, K. Deng, X. Chen, Z. Zou, J. Phys. Chem. C 111 (2007) 2709–2714.
- [38] K.V. Baiju, S. Shukla, K.S. Sandhya, J. James, K.G.K. Warrier, J. Phys. Chem. C 111 (2007) 7612–7622.
- [39] F. Rota, M. Cavassi, D. Niego, R. Gorlani, L. Vianelli, L. Tatti, P. Bruzzi, A. Moroni, I.R. Bellobono, Chemosphere 33 (1996) 2159–2173.